

Ecological species groups and interspecific associations of dominant woody species in a seasonal tropical forest of Laos

Nguyen Van Quy¹, Pham Van Dien², Nguyen Van Thinh¹, Ngo The Long³, Vu Manh¹, Trinh Hien Mai², Le Van Cuong², Pham Thanh Trang², Nguyen Van Hop², Nguyen Thanh Tuan², Khamphet Phomphoumy², Do Phong Luu¹, Nguyen Hong Hai²✉

Quy N.V., Dien P.V., Thinh N.V., Long N.T., Manh V., Mai T.H., Cuong L.V., Trang P.T., Hop N.V., Tuan N.T., Khamphet P., Luu D.P., Hai N.H., 2024. Ecological species groups and interspecific associations of dominant woody species in a seasonal tropical forest of Laos. Ann. For. Res. 67(2): 167-184.

Abstract The study of ecological species groups (ESGs) and interspecific interactions offers important insights into the mechanisms that drive tree species coexistence and enhances our understanding of plant community structure, function, and taxonomy. This study aimed to identify ESGs within a seasonal tropical forest in Phou Khao Khouay National Park, located in central Laos, and to investigate the key environmental factors influencing their distribution. Additionally, the study analysed community stability and interspecific associations among woody species using various statistical techniques, including Variance Ratio (VR), chi-squared tests, Association Coefficients (AC), and percentage of Co-occurrence (PC). To collect data, thirty-two permanent plots, each covering an area of 2500 m² (50 × 50 m), were established. In each plot, floristic data and species abundance were recorded, along with soil samples from two depths (0–10 cm and 10–20 cm) for physicochemical analysis. Vegetation data were classified using Two-Way Indicator Species Analysis (TWINSPAN), and differences between ESGs were evaluated using Analysis of Variance (ANOVA). Indicator Species Analysis (ISA) was applied to identify key species for each ESG, while Canonical Correspondence Analysis (CCA) was employed to explore the relationships between ESGs and environmental variables. The results revealed that the woody plant communities in the study area were relatively stable, exhibiting predominantly positive interspecific associations. Additionally, the distribution of the forty identified ESGs showed a strong correlation with topographical factors – such as elevation, aspect, and slope – and soil properties, including pH, organic matter content, cation exchange capacity, total nitrogen, phosphorus, potassium, and soil texture. These findings underscore the importance of recognizing microhabitats that support the growth and conservation of woody plant species in Phou Khao Khouay National Park, offering valuable insights for future ecological research and conservation initiatives.

Keywords: environmental factors, multivariate analysis, permanent plot, indicator species, similar habitats.

Addresses: ¹Southern Branch of Joint Vietnam-Russia Tropical Science and Technology Research Center, Hochiminh, Vietnam. | ²Vietnam National University of Forestry, Hanoi, Vietnam. | ³Hung Vuong University, Phutho, Vietnam.

✉ **Corresponding Author:** Nguyen Hong Hai (hainh@vnuf.edu.vn).

Manuscript: received March 15, 2024; revised November 25, 2024; accepted December 17, 2024

Introduction

The study of ecological species groups (ESGs) and interspecific interactions has garnered increasing attention due to their pivotal role in elucidating ecological processes across various scientific domains, including ecology, botany, and zoology (Bergeron & Bouchard 1984, Wiegand et al. 2007, Kunwar et al. 2012, Su et al. 2015). ESGs are characterized by the composition and spatial arrangement of species within a defined area (Li et al. 2008). Their identification is based on recognizing species that exhibit similar environmental preferences and occupy comparable ecological niches, thereby forming distinct patterns across landscapes (Archambault et al. 1989). These species groups serve as important indicators of ecosystem health, often consisting of species with a high degree of fidelity to specific niches, and thus are valuable tools for assessing forest ecosystem conditions (Li et al. 2008, Su et al. 2015).

In applied ecology, ESGs are essential for delineating and mapping ecosystem types based on species composition and coverage, providing key insights into ecological classification and habitat management (Abella & Covington 2006, Rad & Shafiei 2010). A more comprehensive understanding of ESGs not only enhances biodiversity monitoring and conservation but also strengthens our grasp of ecosystem resilience and sustainability in the face of environmental change (Daubenmire 1952, Robertson et al. 1988, Burke 2001, Adel et al. 2014).

In ecological communities, species distributions are often uneven, with dominant species being more abundant and conspicuous, while rarer species are less numerous and harder to detect (Grime 1998, Hou et al. 2023). Dominant species typically occupy broader ecological niches, leading to greater overlap with other species and more extensive resource utilization compared to rarer species, which tend to be restricted to specialized niches with limited resources (Song & Zhang 2018, Jiang et al. 2024). Dominant species also exhibit

higher resource acquisition efficiency and greater resilience to extreme environmental conditions, whereas rare species are more likely to be confined to resource-rich habitats (Nguyen et al. 2022, Xue et al. 2023).

Understanding the dynamics among dominant species is crucial for vegetation restoration, particularly in degraded ecosystems such as barren hillsides or disturbed environments (Nguyen et al. 2023c). These species play a key role in enhancing ecosystem resilience and supporting ecological restoration, making them central to efforts aimed at restoring ecological balance in disturbed habitats (Wang & Cui 2023). To ensure the reliability of ecological assessments, it is critical to consider not only the status of dominant tree species but also key environmental factors, such as topography, soil properties, and microclimatic conditions (Adel et al. 2014). Soil structure and nutrient availability are particularly significant, as they directly influence plant growth and community dynamics (Ball et al. 2005, Robertson et al. 1988).

In tropical forests, where soil properties are often heterogeneous, understanding the complex interactions between plant species and their environment is vital for effective ecosystem management and restoration (Crouzeilles et al. 2017, Gatica-Saavedra et al. 2017). Therefore, a comprehensive understanding of dominant species and their interactions with environmental factors is essential for advancing ecological research and implementing successful restoration strategies (Liu et al. 2022).

Interspecific interactions, ranging from facilitation to competition, are fundamental to the stability and dynamics of forest communities, especially across different stages of ecological succession (Fichtner et al. 2017, Pretzsch 2022). These interactions offer critical insights into species relationships and their spatial distribution, thereby improving our understanding of species assembly and community structure evolution (Cavard et al. 2011). By analysing species' habitat

preferences and their interspecific interactions, associations can be classified into positive, negative, or neutral categories (Liu et al. 2014, Nguyen et al. 2023b). Investigating these associations alongside ESG identification is essential for developing effective strategies for vegetation restoration and biodiversity conservation (Adel et al. 2014, Su et al. 2015).

In regions like Laos and the broader Indochinese Peninsula, afforestation efforts have historically focused on monoculture plantations, which often result in simplified stand structures, reduced productivity, and ecological instability (McElwee 2003). Recognizing the limitations of monoculture systems, there has been a shift toward establishing mixed-species forests (McElwee 2016). This transition requires careful selection of appropriate ESGs, a process that is both complex and resource-intensive (Yang et al. 2016). Identifying viable ESGs involves long-term monitoring, making it a time-consuming and costly endeavour (Li et al. 2008). However, recent studies have highlighted the resilience of certain species guilds - particularly stable species pairs - under natural selection, suggesting promising avenues for identifying sustainable ESGs that could enhance afforestation practices and promote integrated forest ecosystem management (Li et al. 2008, Su et al. 2015, Yang et al. 2016).

To effectively classify ESGs across diverse ecosystems, including forests, wetlands, and phytoplankton communities, a combination of traditional and advanced methodologies is essential (Meyer 2006). For example, Abella & Shelburne (2004) employed discriminant analysis and R-mode factor analysis to classify plant species into distinct ESGs, providing a robust framework for ecological classification. Similarly, Li et al. (2008) utilized statistical methods, such as chi-square tests, co-occurrence analysis, and clustering techniques, to study plant associations in the tropical rainforest of Hainan Island, China. Although these approaches have proven

effective in various contexts (Su et al. 2015), ESG research in Laos remains underdeveloped due to logistical challenges and the nascent stage of regional forestry practices. The lack of standardized methodologies for studying interspecific interactions and ESG categorization in Southeast Asia further exacerbates these challenges. Therefore, region-specific research and the application of advanced analytical techniques are urgently needed to address these gaps and deepen our understanding of local ecosystems.

This study, conducted in Phou Khao Khouay (PKK) National Park in central Laos, aims to bridge these gaps by classifying ESGs using multivariate analytical techniques. Specifically, cluster analysis was employed for ESG classification, while ordination methods - such as Indicator Species Analysis (ISA) and Canonical Correspondence Analysis (CCA) - were used to explore species-environment relationships. The study also examined interspecific associations by focusing on dominant tree species across thirty-two permanent plots within the park. Statistical methods, including variance ratio, chi-square tests, association coefficients, and species co-occurrence percentages, were applied to analyse these interactions.

The study addresses five key research questions: (Q1) What are the dominant tree species in the seasonal tropical forest of PKK National Park? (Q2) Are the woody plant communities in the study area stable? (Q3) What interspecific associations exist among the dominant tree species? (Q4) Can the tree species in the study area be classified into distinct ESGs? (Q5) Are there significant variations in environmental factors, such as elevation and soil properties, across the identified ESGs?

The findings from this research provide valuable insights that can inform decision-making in afforestation and forest restoration efforts. Specifically, the study offers a scientific framework for selecting optimal mixed-species

compositions in plantations, facilitating the transition from monoculture to diverse, mixed-species stands. Additionally, the results support enrichment planting in degraded secondary forests and contribute to the development of more effective tropical forest management strategies. By enhancing our understanding of community composition and interspecific interactions, this study provides crucial guidance for habitat conservation and promotes the sustainable management of plant species in the region. Furthermore, it serves as a valuable reference for future research on tropical forest dynamics.

Materials and Methods

Study area

Laos, a landlocked nation in the heart of the Indo-Chinese peninsula, is home to Phou Khao Khouay (PKK) National Park, a protected area since 1993 (Manichanh et al. 2015). Located approximately 40 km northeast of Vientiane, the park covers an area of about 2,000 km² and is situated in a predominantly mountainous region (Lucas et al. 2013). Its geographical coordinates range from 18°14' to 18°32' N and 102°38' to 102°59' E (Fig. 1). To the northwest, PKK National Park is bordered by the Ang Nam Ngum Reservoir, the largest artificial lake in southeast Asia, which marks the park's boundary in that direction.

PKK National Park is home to a variety of forest types, including mixed deciduous forests dominated by Fabaceae, dry evergreen Dipterocarp forests, and monodominant coniferous forests, primarily composed of Pinaceae, at higher elevations (Nguyen et al. 2023a). The park's elevation ranges from 100 m to nearly 1,700 m a.s.l., supporting a variety of ecosystems at different altitudes.

The climate in PKK National Park is characterized by a distinct rainy season from May to October, with an average annual rainfall of approximately 1,769 mm (Chanthalaphone et al. 2020). Precipitation peaks in August, with an average of 494.2 mm. In contrast, the dry season, from November to March, experiences significantly lower rainfall, with February typically recording the lowest monthly average of about 2.5 mm. Temperature fluctuations are also notable: April, the hottest month, sees average temperatures reaching 39°C, while December, the coldest month, averages 10°C. Throughout the year, temperatures range from 16.6°C to 31.8°C during the rainy season and from 16.8°C to 24.6°C during the dry season (Phomphoumy et al. 2023).

The soils in PKK National Park are primarily tropical red to brown soils, including organic acrisols and lithosols, with textures ranging from sandy to sandy-loam and generally low organic matter content (Soukhavong et al. 2013). The combination of diverse ecosystems, varying climatic conditions, and distinct soil types makes PKK National Park an important area for biodiversity conservation and ecological research.

Data collection

This study is based on permanent plots established in 2009 by the Institut de Recherche pour le Développement (IRD), France, in collaboration with the Faculty of Forestry Science at the National University of Laos. A total of thirty-two 50 × 50 m plots were geographically

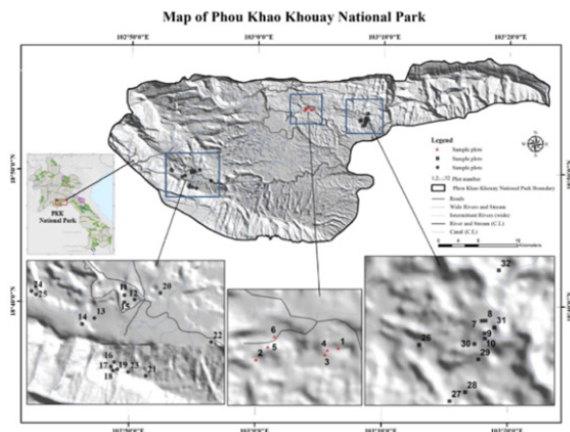


Figure 1 Maps of study region and locations of the study plots.

mapped, with each plot subdivided into 25 subplots of 10 × 10 m. The plots were randomly allocated after stratification by forest type to ensure representativeness across different ecosystems. Within each subplot, all woody tree individuals with a diameter at breast height (DBH) ≥ 5 cm were surveyed. The DBH of each tree was measured, and species were identified by taxonomists from the herbarium of the Faculty of Forestry Science.

In addition to the botanical data, key topographical variables - such as latitude, longitude, aspect, slope, and elevation - were recorded for each plot. These measurements were collected using a compass and a Garmin GPS 60s device.

Five soil pits were excavated to a depth of 20 cm at the four corners and the centre of each plot. Soil samples were then collected from two depth intervals (0–10 cm and 10–20 cm) using a soil drilling sampler with a 5 cm diameter. For each plot, soil samples from the same depth layer were combined in equal volume proportions, air-dried naturally, and stored at room temperature. In total, 64 soil samples were collected (32 plots × 2 depths). Before analysis, all samples were sieved through a 2-mm screen to remove soil fauna, plant roots, and other debris. The samples were then analysed at the soil laboratory of the National Agriculture and Forestry Research Institute (NAFRI) in Laos. A range of soil properties was assessed, including pH, organic matter content (OM), cation exchange capacity (CEC), total nitrogen (TN), total phosphorus (TP), total potassium (TK), and the particle size distribution of sand, clay, and silt.

Data analysis

Importance value index

The importance value index (IVI) functions as a comprehensive metric employed to assess the relative significance of species within a forest community, reflecting their dominance in a specific plot. The magnitude of a species' IVI is directly correlated with its dominance

within the plot. The calculation of IVI followed the formula (Chai et al. 2016, Nguyen et al. 2023b):

$$IVI = (Rd + Ra + Rf) \times 100 / 3 \quad (1)$$

$$\text{Relative dominance (Rd)} = a_i / \sum_{i=1}^S a_i \quad (2)$$

$$\text{Relative abundance (Ra)} = n_i / \sum_{i=1}^S n_i \quad (3)$$

$$\text{Relative frequency (Rf)} = f_i / \sum_{i=1}^S f_i \quad (4)$$

where, S represents the total number of species, a_i corresponds to the basal area of the i th species, n_i signifies the number of individuals of the i th species, and f_i denotes the number of subplots in which the i th species occurred.

Tree species exerting significant ecological influence were identified as those with an IVI ≥ 2% within the study plot (Nguyen et al. 2023c). The collective IVI of tree species surpassing 50% was designated as the dominant species group within the forest communities (Nguyen et al. 2021).

Interspecific association quantification

This study focused on the interspecific associations among dominant woody species in thirty-two study plots. The dominant species were sorted by IVI, and various indices were calculated, including chi-square tests, association coefficients, and the percentage co-occurrence (Jin et al. 2022). The approach involved transforming the original data matrix of S×N (where S represents tree species and N denotes the number of plots) into a binary data matrix in 0-1 format. Subsequently, 2×2 contingency tables were constructed for each species pair, and the values of a, b, c, and d were calculated. Specifically, a represented the number of sample plots containing both species A and B, b signified the number of sample plots with only species B, c denoted the number of plots with only species A, and d accounted for the plots where neither species A nor B was present. The parameter n represented the total number of plots.

In this study, we utilized the variance ratio (VR), introduced by Schluter (1984), to assess the overall association among different species based on their presence or absence in the plots (Chai et al. 2016). The VR equals 1 when

assuming independence, with a VR greater than 1 indicating a positive association among species. Conversely, a VR less than 1 suggests a negative association for the species involved.

The interspecific association between species pairs was assessed using the chi-squared test. Given the non-continuous nature of the samples, the chi-squared value for the study data was determined using Yates' continuity correction formula as follows (Shao & Zhang 2021):

$$\chi^2 = \frac{n(|ad-bc|-0.5n)^2}{(a+b)(a+c)(b+d)(c+d)} \quad (5)$$

A positive association is observed when the product of ad exceeds that of bc , whereas a negative association is evident when the product of ad is less than that of bc . If the calculated chi-squared value surpasses 3.841, it implies a significant association between the species ($0.01 < p < 0.05$). Similarly, a chi-squared value exceeding 6.635 indicates an extremely significant association between pairs of species ($p < 0.01$).

The χ^2 statistic provides only a qualitative assessment of whether the association between species is significant, but it does not offer insight into the strength of these interspecific associations (Liu et al. 2019b). To quantify the strength of the association, the association coefficient (AC) and percentage co-occurrence (PC) are commonly employed. Among these, PC is generally considered more reliable than AC in reflecting the strength of positive associations, as it mitigates the influence of high AC values resulting from large d values or low AC values due to small values. Consequently, a combined analysis of both PC and AC is often conducted to enhance the accuracy and robustness of the results (Liu et al. 2019a).

Association coefficient: the AC is employed to further scrutinize the results obtained through χ^2 and to clarify the strength of interspecific association. Its formula is as follows (Liu et al. 2019a):

$$\text{when } ad \geq bc, AC = \frac{(ad - bc)}{(a + b)(b + d)} \quad (6)$$

$$\text{when } ad < bc \text{ and } d \geq a, AC = \frac{(ad - bc)}{(a + b)(a + c)} \quad (7)$$

$$\text{when } ad < bc \text{ and } d < a, AC = \frac{(ad - bc)}{(a + b)(a + c)} \quad (8)$$

The range of AC values is $[-1, 1]$. As the AC value approaches 1, it indicates a more robust positive association among species pairs; as the AC value approaches -1, it signifies a more pronounced negative association between species; when the AC value is 0, it denotes complete independence among species.

Percentage of co-occurrence: The PC is used to evaluate the degree of positive association between species. Its formula is as follows (Bosun & Shaolin 1985):

$$PC = \frac{a}{(a + b + c)} \quad (9)$$

The range of PC values is $[0, 1]$, with values approaching 1 indicating a stronger positive association between the species.

Data processing was carried out using Excel 2016 software to generate a half-matrix plot based on the calculated χ^2 test values for the associations between various woody species. The AC and PC values obtained were then analysed using R 4.1.3 software and the *vegan* package to generate half-matrix plots (Nguyen et al. 2023c).

Clustering analysis and determination of indicator species

This study utilized the Two-Way Indicator Species Analysis (TWINSPAN) method to classify forest plant communities in PKK National Park, based on species presence/absence data from study plots. TWINSPAN was chosen for its effectiveness in grouping communities with similar species abundance patterns (Adel et al. 2014). The cut-off level for defining 'pseudo-species' followed the default setting of the software. Subsequently, Indicator Species Analysis (ISA) was applied to identify species that were significantly associated with each group. Both TWINSPAN and ISA analyses were conducted using the PC-ORD 5.10 software (McCune & Mefford 2006).

To evaluate the completeness of the sampling and visualize variations in species richness across ESGs, we calculated and plotted sampling coverage using the iNEXT package (Hsieh et al. 2016). For each ESG, sampling coverage was estimated by performing 50 bootstrap replicates, with a 95% confidence interval. This method allows for an assessment of how well the sample reflects the total species richness of ESGs, providing valuable insights into the reliability and stability of the sampling effort (Dar & Parthasarathy 2023).

CANOCO 5 software was used for multivariate analysis to examine the relationship between environmental factors (topographic and soil variables) and the plant community in PKK National Park (Šmilauer & Lepš 2014). Canonical correspondence analysis (CCA), with detrending by segments, was initially performed to assess the gradient length of the first axis and to determine whether linear or unimodal-based numerical methods should be applied (Liu et al. 2012). The significance of the eigenvalues for the first canonical axis was tested using a Monte Carlo permutation test with 499 random permutations. Inter-set correlations from the ordination analysis were used to evaluate the importance of the environmental variables.

The environmental variables considered in the CCA analysis included elevation, slope, slope aspect, soil pH, TN, TP, TK, OM, soil texture (clay, sand, silt), and CEC. Aspect data were transformed using the equation $B' = \cos(45 - B') + 1$ (Beers et al. 1966). The normality of all parameters was assessed using Kolmogorov–Smirnov tests. Differences in the means of environmental variables among groups were analysed using one-way ANOVA,

followed by the Duncan test at the 95% level. All statistical analyses were performed using R 4.1.3 software with the 'vegan' and 'agricolae' packages. Scientific names of species were recalibrated according to Plants of the World Online (<https://powo.science.kew.org>) and World Flora Online (<https://www.worldfloraonline.org>).

Results

Composition of trees species

A total of 5,477 individual trees were recorded across thirty-two permanent plots, representing 194 woody species from sixty-seven plant families (Supplementary materials are available at: <https://github.com/quyforest/data-PKK>). Among these species, ten were identified as dominant, each with an IVI exceeding 2% (Table 1). These ten species collectively comprised 33.2% of the total individuals in the entire species pool. Notably, *H. pierrei* emerged as the most dominant species, with an IVI of 6.53%, while *L. fenestratus* exhibited the lowest IVI among the dominant species, at 2.38%. Despite their prominence in terms of IVI, the individual density of these dominant species remained relatively low, with none exceeding 50 individuals per hectare.

Table 1 Trees species composition, acronyms, and importance values¹.

No.	Species	Acronym	N	D	IVI
1	<i>Hopea pierrei</i> Hance	HOPH	278	35	6.53
2	<i>Schima wallichii</i> (DC.) Korth.	SCWA	266	33	3.51
3	<i>Pinus merkusii</i> Jungh. & de Vriese	PIME	123	15	3.09
4	<i>Alphonsea gaudichaudiana</i> (Baill.) Finet & Gagnep.	ALGA	227	28	2.97
5	<i>Hydnocarpus ilicifolia</i> King	HYIL	224	28	2.87
6	<i>Syzygium cinereum</i> (Kurz) Chantar. & J.Parn.	SYCI	187	23	2.75
7	<i>Gironniera nervosa</i> Planch.	GINE	162	20	2.56
8	<i>Xanthophyllum lanceatum</i> J.J.Sm.	XALA	74	9	2.46
9	<i>Syzygium syzygioides</i> (Miq.) Merr. & L.M.Perry	SYSY	124	16	2.45
10	<i>Lithocarpus fenestratus</i> Rehder	LIFE	154	19	2.38
11	Ten dominant species	----	1,819	227	31.57
12	Others (184 species)	----	3,658	457	68.43
13	All (194 species)	----	5,477	685	100

¹N represents the number of individuals, D signifies individual density per hectare (individuals ha⁻¹), and IVI stands for Importance Value Index, expressed as a percentage (%).

Interspecific associations among species

Overall interspecific association

The calculated VR value of 4.68 for the 194

also satisfied the criteria set by the AC and PC tests, reinforcing the robustness of these associations.

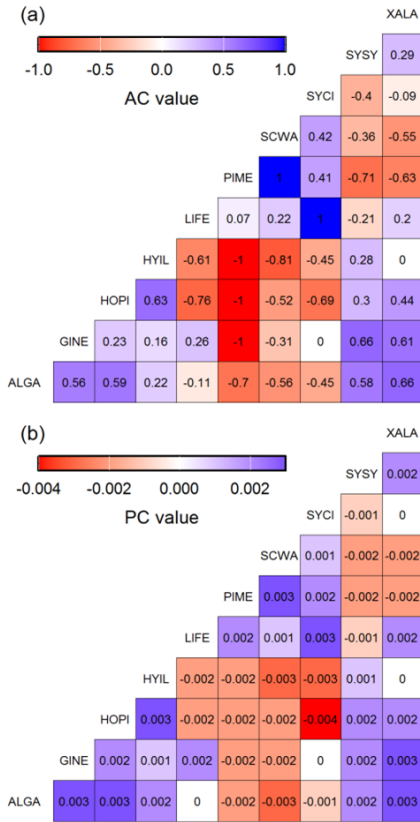


Figure 3 AC and PC results of dominant species in the study area.

Positive associations are indicated in blue, while negative associations are shown in red. Species acronyms correspond to those in Table 1.

Division of ecological species groups

The TWINSpan analysis identified four distinct ESGs based on indicator species (Fig. 4). The first group, which included seven plots (01, 02, 09, 10, 29, 31, 32), was characterized by the presence of three indicator species: *A. gaudichaudiana*, *A. polystachya*, and *C. iners*. The second group, consisting of eight plots (05, 06, 11, 12, 15, 16, 20, 22), was defined by the presence of three species: *L. fenestratus*, *S. cinereum*, and *S. wallichii*. The third group,

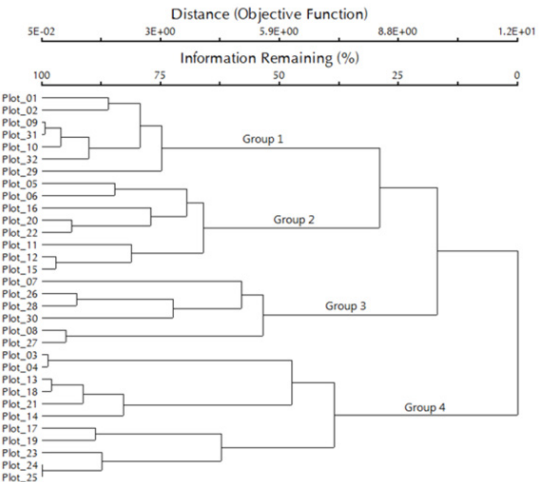


Figure 4 Diagram of TWINSpan analyses on thirty-two plots in the study area.

which included six plots (07, 08, 26, 27, 28, 30), was associated with three indicator species: *H. ilicifolia*, *S. siamensis*, and *C. glabrum*. Finally, the fourth group, comprising 11 plots (03, 04, 13, 14, 17, 18, 19, 21, 23, 24, 25), was characterized by three indicator species: *G. nervosa*, *A. gomezianus*, and *B. macrostachya*.

The sample coverage across the four ESGs was nearly identical: Group 1 = 98.25%, Group 2 = 97.88%, Group 3 = 98.88%, and Group 4 = 99.34%, demonstrating that the samples were almost equally complete (Fig. 5). Extrapolation

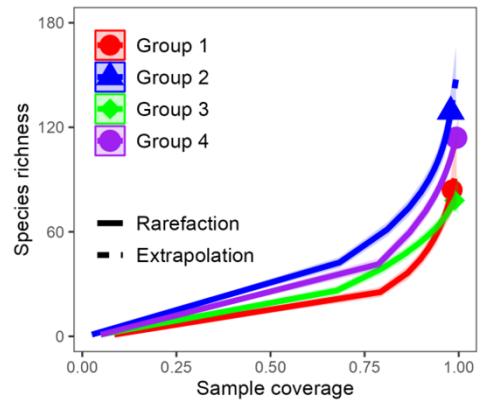


Figure 5 Coverage-based rarefaction and extrapolation curves with 95% confidence intervals, comparing plant species richness across four ecological groups in the seasonal tropical forest of PKK National Park. The solid lines represent the reference samples, while the extrapolation phase is indicated by the dashed lines.

across the communities resulted in negligible increases in sample coverage—specifically, 0.00197%, 0.00196%, 0.00198%, and 0.00199% for Groups 1, 2, 3, and 4, respectively. These findings from coverage estimation indicated that the samples for each ESG had been almost complete, and further sampling would not significantly alter these results.

The results obtained from the CCA showed the impact of soil and topography variables on the ESGs. Particularly, the first and second axes, possessing eigenvalues of 0.529 and 0.397 respectively, collectively explained 92.6% of the total variance (Fig. 6). This finding indicated that TK, OM, and pH were pivotal determinants within the first group, while TP and clay played significant roles in the second group. Additionally, aspect and slope emerged as influential factors in the third group, whereas elevation, sand, TN, CEC, and silt contributed to distinguishing the fourth group.

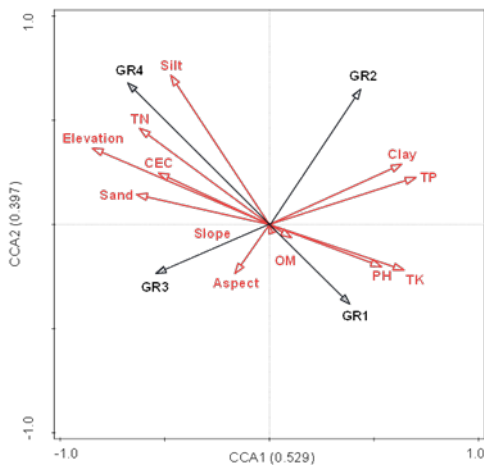


Figure 6 The diagram illustrates the results of a CCA examining the correlation between species groups and environmental factors in the studied area. Abbreviations are employed for ecological species groups, with GR denoting a specific group.

The ANOVA analysis revealed significant differences in the associations of ESGs with soil and topographic variables (Fig. 7). Specifically, Group 1 was characterized by significantly higher levels of soil pH, OM, TK, and sand content (Fig. 7d, e, i, j). These findings

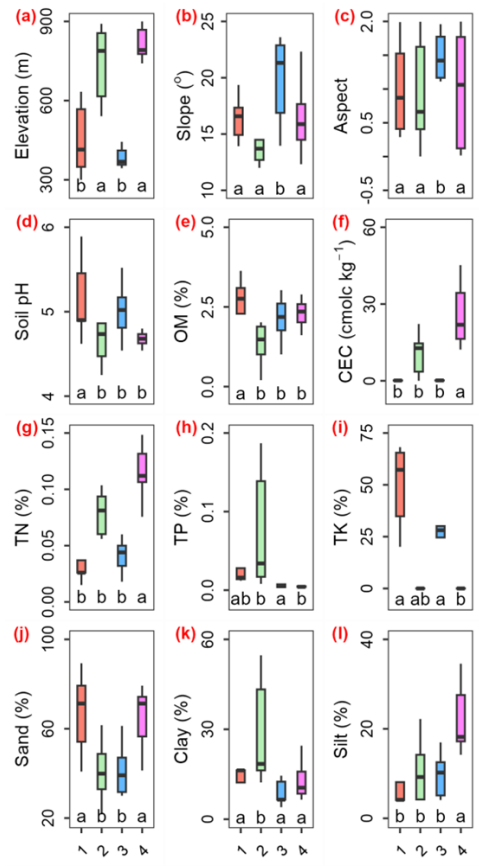


Figure 7 Differences in environmental variables across ecological species groups.

suggest that the species in Group 1 are adapted to, or preferentially inhabit, environments with more alkaline soils, higher nutrient availability, and coarser-textured soils, which are typically associated with better drainage.

In contrast, Group 2 exhibited significantly higher clay content and TP, indicating a preference for soils with finer textures and higher phosphorus concentrations (Fig. 7h, k). This suggests that species in Group 2 favor habitats with higher moisture retention and enhanced nutrient availability - conditions typically found in soils with lower drainage capacities. Group 3, on the other hand, showed the highest values for slope and aspect, indicating a preference for more variable

or sloped terrains (Fig. 7b, c). This pattern reflects an adaptation to specific topographical features, such as water drainage patterns, sunlight exposure, and temperature variations, all of which likely influence the suitability of habitats for the species in this group.

Finally, Group 4 was associated with elevated levels of silt, sand, CEC, TN, and elevation (Fig. 7a, f, g, l, j). These results suggest that species in Group 4 are linked to soils with a balanced texture (silt and sand), higher nutrient retention capacity (as indicated by CEC and TN), and are typically found at higher elevations, where environmental conditions such as temperature and humidity differ from those at lower altitudes.

Discussion

Overall interspecific association of tree species

The process of community succession involves dynamic changes in species interactions, which are crucial for understanding the mechanisms that maintain community stability and biodiversity (Chai et al. 2016, Detto et al. 2022). Shuai & Wang (2023) emphasize that environmental filtering and interspecific competition play central roles in shaping species' functional traits and ecological strategies during succession. In the early stages of succession, these processes often result in competitive exclusion or niche differentiation (Buche et al. 2022). However, as succession progresses, species traits tend to complement one another, leading to more mutualistic or neutral interactions (Cassini 2020, Falk et al. 2022, Löffler & Pape 2020). Our findings regarding the overall association among 194 woody species in the seasonal tropical forest of PKK National Park, Laos, align with this model, as evidenced by the overall positive association observed among species. This suggests that the woody plant communities in the study area are entering a more stable phase, characterized by cooperative or mutually

beneficial interspecific relationships.

The shift in species interactions indicates that the community is approaching equilibrium, where species are better adapted to coexist (Chen et al. 2018). This transition reduces the likelihood of competitive exclusion and enhances the community's resilience (Li et al. 2020). The positive association observed among the 194 woody species in the study area points to a trend toward more stable, potentially symbiotic relationships as succession progresses. These interactions suggest that the woody plant communities in PKK National Park are nearing, or have already reached, the final stage of ecological succession, where the competitive dynamics of earlier stages are replaced by cooperative associations.

It is important to note that this equilibrium does not imply the complete absence of competition (Aarssen 1983). Rather, it signifies a balance between competitive interactions and cooperative relationships, allowing species to coexist without compromising the community's overall functionality (Shmida & Ellner 1984). Our results offer a deeper understanding of community dynamics by illustrating how interspecific relationships evolve from competitive dominance to cooperative stability as succession unfolds in the woody plant communities of PKK National Park. These findings support the idea that as communities mature, their internal interactions become more integrated, leading to a more stable and resilient ecological system (Jin et al. 2022).

Different letters below the boxplot (Fig. 7) indicate significant differences between various species groups in the study area ($p < 0.05$).

Interspecific associations between dominant species

Interspecific associations provide valuable insights into the interactions between species across diverse habitats, offering a deeper understanding of their capacity to adapt to varying environmental conditions (Liu et al.

2014, Liu et al. 2019b). This study focused on analysing interspecific associations based on species pairs, revealing that dominant species exhibited both positive and negative associations in roughly equal proportions. However, a clear disparity emerged in the frequency of significant positive versus negative associations. Specifically, the number of species pairs with significant positive associations exceeded those with significant negative associations, suggesting a general trend of positive interspecific relationships among the 194 species in the study area.

Positive associations between species pairs typically indicate similar or shared environmental resource requirements, often pointing to potential mutualistic or symbiotic relationships (Nguyen et al. 2022). In contrast, negative associations generally reflect species' ability to adapt to environmental heterogeneity, where variations in biological traits can lead to competitive exclusion and niche differentiation (Phuong et al. 2022). The size of the sampling plot is a critical factor influencing the nature of interspecific associations (Nguyen et al. 2023c). Larger habitats tend to show predominantly positive associations, while smaller habitats are more likely to exhibit negative associations (Shuai & Wang 2023). Empirical studies in tropical forests suggest that optimal plot sizes range from 400 m² to 900 m² (Chen et al. 2018). In this study, the use of a larger plot size of 2500 m² enhanced the reliability of the findings by providing a broader, more representative analysis.

Although the plot size and number of plots in our study were comparable to those of Pham et al. (2021) and Huong & Cuong (2022), our results exhibited notable differences. Specifically, the proportion of species pairs showing no significant interspecific associations was considerably lower in our study. One key difference between our study and those of Pham et al. (2021) and Huong & Cuong (2022) lies in the statistical methods used. While their studies employed an

unadjusted chi-square test, our analysis applied Yates's corrected chi-square test. Yates's correction for continuity enhances the accuracy of the chi-square test, particularly in cases with small sample sizes in 2×2 contingency tables (Adler 1951). The unadjusted chi-square test is typically less conservative and can increase the likelihood of rejecting the null hypothesis (Haber 1982). In contrast, Yates's correction is more conservative, reducing the risk of Type I errors and thus increasing the reliability of hypothesis testing (Li et al. 2020). This methodological refinement strengthened the robustness of our findings and aligns with contemporary statistical advancements (Nguyen et al. 2023c).

To further verify the robustness of our results, we employed both the AC and PC tests, which yielded findings consistent with the initial chi-square analysis of interspecific associations. The outcomes from these additional tests confirmed the reliability of our preliminary results. Specifically, the nine species pairs showing positive associations and the seven exhibiting negative associations, as identified through the chi-square test, all exhibited AC values greater than 0.5 and PC values exceeding 0.002. Consequently, the 16 species pairs with significant associations in the chi-square analysis also met the criteria established by the AC and PC tests, further reinforcing the strength and reliability of these associations.

Division of ecological species groups

In our study, TWINSpan analysis was applied to classify community types within PKK National Park. It was previously demonstrated by Dar & Parthasarathy (2023) that TWINSpan provides a more accurate representation of study plot distribution and species composition compared to other clustering methods, such as hierarchical clustering, k-means, or model-based clustering. A key strength of TWINSpan lies in its ability to identify indicator species that are strongly associated with specific environmental conditions or

habitat types, making it a valuable tool in ecological studies focused on species-environment interactions (Hill 1979). Unlike other methods, TWINSpan incorporates these indicator species into the clustering process, thereby enhancing the ecological relevance of the resulting groups (Clare 2000, Hugh G. Gauch & Whittaker 1981). Additionally, its binary division approach facilitates clear, dichotomous classification, producing well-defined, easily interpretable groups that effectively capture ecological gradients or differentiate distinct community types (Cui et al. 2009, Marshall & Elliott 1997). This feature is particularly beneficial in ecological data analysis, where identifying discrete community units is crucial for effective ecological management and conservation (Song & Zhang 2018). The incorporation of conservation strategies that utilize spatial patterns of species richness and diversity - such as prioritizing conservation areas based on key floristic groups - can improve forest policy and management effectiveness (Khan et al. 2011).

The application of TWINSpan in this study enabled the identification and characterization of vegetation groups, facilitating an interpretation of their diversity and species distribution patterns (Siebert 2012). However, challenges arose due to the discontinuous distribution of communities, as effective classification depends on relatively stable species composition, structure, and habitat occupation (Li et al. 2020, van der Maarel & Sykes 1993). Differences in microhabitats, stages of community development, and species interactions within similar habitats were significant in distinguishing community types in tropical forests (Rees et al. 2001). The classification of species groups and communities in our study primarily relied on assessments of species abundance, presence/absence, and environmental factors, revealing significant correlations between species composition and various topographical and

soil characteristics, as demonstrated by CCA analysis. These findings underscore the considerable influence of environmental factors on species distribution within the study area.

Four distinct ESGs were identified within the woody plant communities in PKK National Park. Vilches et al. (2013) emphasized that ESGs, when analysed in conjunction with physiographic, microclimatic, and soil variables, are effective tools for delineating ecosystems across multiple landscape scales. Our analysis confirmed that the distribution of these ecological groups was closely linked to the topographical and soil properties of the sampled plots. Through the use of CCA and ANOVA, clear relationships were found between ESG distribution and environmental factors. Specifically, aspect and slope were significant determinants of the third ESG, while elevation had a substantial influence on the fourth ESG. Soil variables such as pH, OM, TK, and sand content were key factors shaping the first ESG, whereas clay content and TP were central to the second ESG. The composition of the third ESG was most strongly influenced by silt content, sand, TN, and CEC.

Our findings align with previous research. For instance, Olivero & Hix (1998) emphasized the role of aspect in shaping ground flora distribution. The geographical distribution of plant species is influenced by factors such as water availability, light, and soil temperature (Buri et al. 2017, Huang et al. 2021). Topography indirectly affects these resources by modifying solar radiation, which in turn impacts temperature, soil moisture, humidity, vapor pressure deficits, and transpiration - all crucial for plant survival, growth, and distribution (Small & McCarthy 2005). Similarly, Dar & Parthasarathy (2023) suggested that altitude, along with slope and aspect, affects solar azimuth, hydrology, climate, and soil properties, influencing vegetation patterns through the creation of

microenvironments.

Our study further emphasized the significant role of soil variables in environmental differentiation, ranking second only to topography. Through comprehensive analyses, including CCA and ANOVA, we observed that soil nutrient content plays a critical role in ESG distribution. This finding aligns with prior research demonstrating the impact of soil nutrients on plant community composition in tropical forests. For example, Mataji et al. (2009) examined soil characteristics within the Rusco-Fagetum ecological group in Iran, finding that clay soils influence plant community formation. Soil nutrients are essential for plant nutrition, as their concentration and availability largely determine soil fertility and site productivity (El-Ramady et al. 2014). Furthermore, Amorin & Batalha (2007) identified phosphorus as a key driver of plant community dynamics in Brazil, further underscoring the importance of soil nutrients in shaping ecological groupings.

Conclusions

This study provides valuable insights into the stability and interspecific associations of plant communities in the natural forests of PKK National Park, Lao PDR. By analysing data from thirty-two permanent plots and documenting 5,477 individual trees across 194 species, we identified the key characteristics of tropical forests in the region.

The study revealed that plant communities are in a stable phase, with positive associations among most species and specific ecological groupings identified through indicator species analysis. Notably, environmental factors such as topography and soil properties were found to significantly influence plant distribution and the formation of ESGs.

Furthermore, the identification of dominant species and their interrelationships, particularly through the analysis of 45 species pairs, highlights the complex dynamics within the forest ecosystem. These findings not only enhance our understanding of plant community

structure but also provide a basis for future monitoring efforts, which could serve as an early warning system for biodiversity conservation.

The results underscore the importance of continuous observation and the potential for using ecological indicators to inform conservation strategies.

Conflict of interest

The authors declare that they have no conflict of interest.

References

- Aarssen L.W., 1983. Ecological combining ability and competitive combining ability in plants: toward a general evolutionary theory of coexistence in systems of competition. *The American Naturalist* 122(6): 707-731. <https://www.journals.uchicago.edu/doi/10.1086/284167>
- Abella S.R., Covington W.W., 2006. Vegetation-environment relationships and ecological species groups of an Arizona *Pinus ponderosa* landscape, USA. *Plant Ecology* 185: 255-268. <https://doi.org/10.1007/s11258-006-9102-y>
- Abella S.R., Shelburne V.B., 2004. Ecological species groups of South Carolina's Jocassee Gorges, southern Appalachian Mountains. *Journal of the Torrey Botanical Society* 131(3): 220-231. <https://doi.org/10.2307/4126952>
- Adel M.N., Pourbabaei H., Dey D.C., 2014. Ecological species group - Environmental factors relationships in unharvested beech forests in the north of Iran. *Ecological engineering* 69: 1-7. <https://doi.org/10.1016/j.ecoleng.2014.03.008>
- Adler F., 1951. Yates' correction and the statisticians. *Journal of the American Statistical Association* 46(256): 490-501. <https://doi.org/10.2307/2280398>
- Amorin P., Batalha M., 2007. Soil vegetation relationship in hyper seasonal cerrado and wet grassland in Emas national park (central Brazil). *Acta Oecol* 32(3): 319-27. <https://doi.org/10.1016/j.actao.2007.06.003>
- Archambault L., Barnes B.V., Witter J.A., 1989. Ecological species groups of oak ecosystems of southeastern Michigan. *Forest Science* 35(4): 1058-1074. <https://doi.org/10.1093/forestscience/35.4.1058>
- Ball B., Bingham I., Rees R., Watson C., Litterick A., 2005. The role of crop rotations in determining soil structure and crop growth conditions. *Canadian Journal of Soil Science* 85(5): 557-577. <https://doi.org/10.4141/S04-078>
- Beers T.W., Dress P.E., Wensel L.C., 1966. Notes and Observations: Aspect Transformation in Site Productivity Research. *Journal of Forestry* 64(10): 691-692. <https://doi.org/10.1093/jof/64.10.691>

- Bergeron Y., Bouchard A., 1984. Use of ecological groups in analysis and classification of plant communities in a section of western Quebec. *Vegetatio* 56: 45-63. <https://doi.org/10.1007/BF00036136>
- Bosun W., Shaolin P., 1985. Studies on the measuring techniques of interspecific association of lower-subtropical evergreen-broadleaved forests. *Chinese Journal of Plant Ecology* 9(4): 274-285. <https://www.plant-ecology.com/EN/>
- Buche L., Spaak J.W., Jarillo J., De Laender F., 2022. Niche differences, not fitness differences, explain predicted coexistence across ecological groups. *Journal of Ecology* 110(11): 2785-2796. <https://doi.org/10.1111/1365-2745.13992>
- Buri A., Cianfrani C., Pinto-Figueroa E., Yashiro E., Spangenberg J.E., Adatte T., Verrecchia E., Guisan A., Pradervand J.-N., 2017. Soil factors improve predictions of plant species distribution in a mountain environment. *Progress in Physical Geography* 41(6): 703-722. <https://doi.org/10.1177/0309133317738162>
- Burke A., 2001. Classification and ordination of plant communities of the Naukluft Mountains, Namibia. *Journal of Vegetation Science* 12(1): 53-60. <https://doi.org/10.1111/j.1654-1103.2001.tb02616.x>
- Cassini M.H., 2020. A review of the critics of invasion biology. *Biological Reviews* 95(5): 1467-1478.
- Cavard X., Bergeron Y., Chen H.Y., Paré D., Laganière J., Brassard, B., 2011. Competition and facilitation between tree species change with stand development. *Oikos* 120(11): 1683-1695. <https://www.jstor.org/stable/41315984>
- Chai Z.Z., Sun C.L., Wang D.X., Liu W.Z., 2016. Interspecific associations of dominant tree populations in a virgin old-growth oak forest in the Qinling Mountains, China. *Botanical studies* 57: 1-13. <https://doi.org/10.1186/s40529-016-0139-5>
- Chanthaphone K., Truong L.X., Nguyen V.T., 2020. Structural characteristics and plant species diversity of some natural forest types at Phou Khao Khouay national park, Lao people's democratic republic. *Journal of Forestry Science and Technology* 9(2020): 053-063.
- Chen Q., Chen J., Zhong J.J., Ji L.T., Kang B., 2018. Interspecific association and functional group classification of the dominant populations in shrub layer in secondary forest of *Pinus tabuliformis* in Qinling Mountain, China. *The Journal of Applied Ecology* 29(6): 1736-1744. <https://doi.org/10.13287/j.1001-9332.201806.004>
- Clare T., 2000. An assessment of the potential of the TWINSpan program of multi-variate analysis to contribute to the classification and management of village landscapes, with reference to historical features. *Landscape Research* 25(1): 117-139. <https://doi.org/10.1080/014263900113208>
- Crouzeilles R., Ferreira M.S., Chazdon R.L., Lindenmayer D.B., Sansevero J.B., Monteiro L., Iribarrem A., Latawiec A.E., Strassburg B.B., 2017. Ecological restoration success is higher for natural regeneration than for active restoration in tropical forests. *Science advances* 3(11): e1701345. <https://doi.org/10.1126/sciadv.1701345>
- Cui B., Zhai H., Dong S., Chen B., Liu S., 2009. Multivariate analysis of the effects of edaphic and topographical factors on plant distribution in the Yilong lake basin of Yun-Gui Plateau, China. *Canadian Journal of Plant Science* 89(1): 211-221. <https://doi.org/10.4141/CJPS08013>
- Dar A.A., Parthasarathy N., 2023. Herb stratum diversity and community structure in Gurez valley of Kashmir Himalaya: application of multivariate techniques in community analyses. *Geology, Ecology, and Landscapes*: 1-17. <https://doi.org/10.1080/24749508.2023.2216532>
- Daubenmire R., 1952. Forest vegetation of Northern Idaho and adjacent Washington, and its bearing on concepts of vegetation classification. *Ecological Monographs* 22(4): 301-330. <https://doi.org/10.2307/1948472>
- Detto M., Levine J.M., Pacala S.W., 2022. Maintenance of high diversity in mechanistic forest dynamics models of competition for light. *Ecological Monographs* 92(2): e1500. <https://doi.org/10.1002/ecm.1500>
- El-Ramady H.R., Alshaal T., Shehata S., Domokos-Szabolcsy É., Elhawat N., Prokisch J., Fári M., Marton L., 2014. Plant nutrition: from liquid medium to micro-farm. *Sustainable Agriculture Reviews* 14: Agroecology and Global Change: 449-508. https://doi.org/10.1007/978-3-319-06016-3_12
- Falk D.A., Van Mantgem P.J., Keeley J.E., Gregg R.M., Guiterman C.H., Tepley A.J., Young D.J., Marshall L.A., 2022. Mechanisms of forest resilience. *Forest Ecology and Management* 512: 120129. <https://doi.org/10.1016/j.foreco.2022.120129>
- Fichtner A., Härdtle W., Li Y., Bruelheide H., Kunz M., Von Oheimb G., 2017. From competition to facilitation: how tree species respond to neighbourhood diversity. *Ecology Letters* 20(7): 892-900. <https://doi.org/10.1111/ele.12786>
- Gatica-Saavedra P., Echeverría C., Nelson C.R., 2017. Ecological indicators for assessing ecological success of forest restoration: a world review. *Restoration Ecology* 25(6): 850-857. <https://doi.org/10.1111/rec.12586>
- Grime J.P., 1998. Benefits of plant diversity to ecosystems: immediate, filter and founder effects. *Journal of Ecology* 86(6): 902-910. <https://doi.org/10.1046/j.1365-2745.1998.00306.x>
- Haber M., 1982. The continuity correction and statistical testing. *International Statistical Review/Revue Internationale de Statistique* 50(2): 135-144. <https://doi.org/10.2307/1402597>
- Hill M.O., 1979. A FORTRAN program for arranging multivariate data in an ordered two-way table by classification of the individuals and attributes. New York, United States: Cornell University.
- Hou G., Shi P., Zhou T., Sun J., Zong N., Song M., Zhang X., 2023. Dominant species play a leading role in shaping community stability in the northern Tibetan

- grasslands. *Journal of Plant Ecology* 16(3): rtac110. <https://doi.org/10.1093/jpe/rtac110>
- Hsieh T.C., Ma K.H., Chao A., 2016. iNEXT: An R package for rarefaction and extrapolation of species diversity (Hill numbers). *Methods in Ecology and Evolution* 7: 1451-1456. <https://doi.org/10.1111/2041-210X.12613>
- Huang E., Chen Y., Fang M., Zheng Y., Yu S., 2021. Environmental drivers of plant distributions at global and regional scales. *Global Ecology and Biogeography* 30(3): 697-709. <https://doi.org/10.1111/geb.13251>
- Hugh G. Gauch J., Whittaker R.H., 1981. Hierarchical classification of community data. *The Journal of Ecology* 69(2): 537-557. <https://doi.org/10.2307/2259682>
- Huong P.V., Cuong L.V., 2022. The ecological interaction between endangered, precious and rare woody species in rich forest community of Tanphu protection forest, Vietnam. *Biodiversitas* 23(12): 6119-6127. <https://doi.org/10.13057/biodiv/d231205>
- Jiang D.D., Luo Y.H., Lin J.Y., He Q.P., Qin L., Ling Y.M., 2024. Spatial distribution pattern and correlation of dominant species of evergreen broad-leaved forest in Shiwandashan mountain. *Journal of Central South University of Forestry & Technology* 44(1): 151-161. <https://link.oversea.cnki.net/doi/10.14067/j.cnki.1673-923x.2024.01.015>
- Jin S.S., Zhang Y.Y., Zhou M.L., Dong X.M., Chang C.H., Wang T., Yan D.F., 2022. Interspecific association and community stability of tree species in natural secondary forests at different altitude gradients in the southern Taihang Mountains. *Forests* 13(3): 373. <https://doi.org/10.3390/f13030373>
- Khan S.M., Harper D., Page S., Ahmad H., 2011. Species and community diversity of vascular flora along environmental gradient in Naran Valley: A multivariate approach through indicator species analysis. *Pak. J. Bot* 43(5): 2337-2346. <http://www.pakbs.org/pjbot/papers/1524568432.pdf>
- Kunwar R.M., Mahat L., Sharma L.N., Shrestha K.P., Kominee H., Bussmann R.W., 2012. Underutilized plant species in far west Nepal. *Journal of Mountain Science* 9: 589-600. <https://doi.org/10.1007/s11629-012-2315-8>
- Li S.F., Lang X.D., Huang X.B., Wang Y.H., Liu W.D., Xu C.H., Su J.R., 2020. Association classification of a 30 hm² dynamics plot in the monsoon broad-leaved evergreen forest in Pu'er, Yunnan, China. *Chinese Journal of Plant Ecology* 44(3): 236. <https://doi.org/10.17521/cjpe.2019.0268>
- Li Y., Xu H., Chen D., Luo T., Mo J., Luo W., Chen H., Jiang Z., 2008. Division of ecological species groups and functional groups based on interspecific association—a case study of the tree layer in the tropical lowland rainforest of Jianfenling in Hainan Island, China. *Frontiers of Forestry in China* 3: 407-415. <https://doi.org/10.1007/s11461-008-0049-0>
- Liu L.G., Zhu Z.Q., Chen X., Liu S.J., 2022. Ecological niche and interspecific association of dominant species in the initial stage of vegetation restoration in Zhongba village of Xide county. *Journal of West China Forestry Science* 51(1): 110-117. <https://m.fx361.com/news/2022/0219/12462391.html>
- Liu X., Zhang W., Yang F., Zhou X., Liu Z., Qu F., Lian S., Wang C., Tang X., 2012. Changes in vegetation–environment relationships over long-term natural restoration process in Middle Taihang Mountain of North China. *Ecological Engineering* 49: 193-200. <https://doi.org/10.1016/j.ecoleng.2012.06.040>
- Liu Y., Li F., Jin G., 2014. Spatial patterns and associations of four species in an old-growth temperate forest. *Journal of Plant Interactions* 9(1): 745-753. <https://doi.org/10.1080/17429145.2014.925146>
- Liu Z., Bai Y., Jiang C., Wang S., Meng J., 2019a. Species association of the dominant tree species in an old-growth forest and their implications for silvicultural practices in western Hunan Province, China. *Austrian Journal of Forest Science* 3: 219–248. https://www.forestsience.at/content/dam/holz/forest-science/2019/03/CB1903_Art3.pdf
- Liu Z., Zhu Y., Wang J., Ma W., Meng J., 2019b. Species association of the dominant tree species in an old-growth forest and implications for enrichment planting for the restoration of natural degraded forest in subtropical China. *Forests* 10(11): 957. <https://doi.org/10.3390/f10110957>
- Löffler J., Pape R., 2020. Thermal niche predictors of alpine plant species. *Ecology* 101(1): e02891. <https://doi.org/10.1002/ecy.2891>
- Lucas C., Nanthavong K., Millet J., 2013. Environmental and human influence on forest composition, structure and diversity in Laos. *Journal of Tropical Forest Science* 25: 410-420. <https://www.jstor.org/stable/23617243>
- Manichanh S., Jérôme M., Andreas H., Khamseng N., Rhett D.H., 2015. Using Plant Functional Traits and Phylogenies to Understand Patterns of Plant Community Assembly in a Seasonal Tropical Forest in Lao PDR. *Plos One* 10(6): e0130151. <https://doi.org/10.1371/journal.pone.0130151>
- Marshall S., Elliott M., 1997. A comparison of univariate and multivariate numerical and graphical techniques for determining inter- and intraspecific feeding relationships in estuarine fish. *Journal of Fish Biology* 51(3): 526-545. <https://doi.org/10.1111/j.1095-8649.1997.tb01510.x>
- Mataji A., Zahedi G., Asri Y., 2009. Vegetation analysis based on plant associations and soil properties in natural forests. *Iranian journal of forest and poplar research* 17(1): 98-85. https://ijfpr.areeo.ac.ir/article_107891_7ab8b28c3d74a5422f03459100975a05.pdf?lang=en
- Mccune B., Mefford M., 2006. *Multivariate analysis of ecological data, version 5.10*. Gleneden Beach, Oregon, USA: MjM software design.
- Mcelwee P.D., 2003. 'Lost worlds' or 'lost causes'? Biodiversity conservation, forest management, and rural life in Vietnam. New Haven, Connecticut, USA: Yale University.

- Mcelwee P.D., 2016. *Forests are gold: Trees, people, and environmental rule in Vietnam*. Seattle, WA, United States of America: University of Washington Press.
- Meyer H.A., 2006. Interspecific association and substrate specificity in tardigrades from Florida, southeastern United States. *Hydrobiologia* 558: 129-132. <https://doi.org/10.1007/s10750-005-1411-y>
- Nguyen H.H., Khamphet P., Nguyen V.Q., 2023a. Nearest Neighbor Patterns of Dominant Tree Species in Tropical Forests, Phou Khao Khouay National Park, Laos. *Journal of Forestry Science and Technology* 15: 16-26. <https://doi.org/10.55250/jo.vnuf.2023.15.016-026>
- Nguyen V.Q., Kang Y.X., Islam A., Li M., Nguyen T.T., Nguyen V.Q., Nguyen V.H., 2022. Spatial distribution and association patterns of *Hopea pierrei* Hance and other tree species in the Phu Quoc Island evergreen broadleaved forest of Vietnam. *Applied Ecology and Environmental Research* 20(2): 1911-1933. http://dx.doi.org/10.15666/aecer/2002_19111933
- Nguyen V.Q., Kang Y.X., Khot C., Nguyen V.H., Nguyen T.T., 2021. Spatial distribution and interspecific association patterns between *Shorea roxburghii* G. Don and other tree species in a South Vietnam evergreen forest. *Applied Ecology and Environmental Research* 19(6): 4665-4681. http://dx.doi.org/10.15666/aecer/1906_46654681
- Nguyen V.Q., Le V.C., Pham T.H., Nguyen H.H., 2023b. Interspecific association of dominant tree species in an evergreen broadleaved forest in Phu Quoc National Park, Vietnam. *Journal of forestry science and technology* 8(2): 87-96. <https://doi.org/10.55250/jo.vnuf.8.2.2023.087-096>
- Nguyen V.Q., Pham V.D., Bui T.D., Nguyen H.H., 2023c. Niche and Interspecific Association of Dominant Tree Species in an Evergreen Broadleaved Forest in Southern Vietnam. *Moscow University Biological Sciences Bulletin* 78(2): 89-99. <https://doi.org/10.3103/S0096392523020062>
- Olivero A.M., Hix D.M., 1998. Influence of aspect and stand age on ground flora of southeastern Ohio forest ecosystems. *Plant Ecology* 139: 177-187. <https://doi.org/10.1023/A:1009758501201>
- Pham V.H., Le H.V., Nguyen T.H., Duong A.T., Kieu P.A., Pham T.L., 2021. Status quo of ecological group between endangered, valuable and rare woody species in rich forest state at Tanphu protection forest - Dong Nai. *VNJFST* 6(1): 60-68.
- Phomphoumy K., Hien C.T.T., Hai N.H., 2023. The relationships of taxonomic and structural attributes on above ground carbon biomass of tropical dry forests in Phou Khao Khouay national park, Laos. *Journal of Forestry Science and Technology* (15): 027-037.
- Phuong P.M., Quy N.V., Huong N.T.M., Tuan N.T., Ha P.T., 2022. Spatial Distribution Patterns and Associations of Woody Plant Species in the Evergreen Broadleaved Forests in Central Vietnam. *Biology Bulletin* 49(5): 369-380. <http://dx.doi.org/10.1134/S1062359022050132>
- Pretzsch H., 2022. Facilitation and competition reduction in tree species mixtures in Central Europe: Consequences for growth modeling and forest management. *Ecological Modelling* 464: 109812. <https://doi.org/10.1016/j.ecolmodel.2021.109812>
- Rad J.E., Shafiei A.B., 2010. The distribution of ecological species groups in Fagetum communities of Caspian forests: determination of effective environmental factors. *Flora-Morphology, Distribution, Functional Ecology of Plants* 205(11): 721-727. <https://doi.org/10.1016/j.flora.2010.04.015>
- Rees M., Condit R., Crawley M., Pacala S., Tilman D., 2001. Long-Term Studies of Vegetation Dynamics. *Science* 293(5530): 650-655. <https://doi.org/10.1126/science.1062586>
- Robertson G.P., Hutson M.A., Evans F.C., Tiedje J.M., 1988. Spatial variability in a successional plant community: patterns of nitrogen availability. *Ecology* 69(5): 1517-1524. <https://doi.org/10.2307/1941649>
- Schluter D., 1984. A variance test for detecting species associations, with some example applications. *Ecology* 65: 998-1005. <https://doi.org/10.2307/1938071>
- Shao L.Y., Zhang G.F., 2021. Niche and interspecific association of dominant tree populations of *Zelkova schneideriana* communities in eastern China. *Botanical Sciences* 99(4): 823-833. <https://doi.org/10.17129/botsci.2809>
- Shmida A., Ellner S., 1984. Coexistence of plant species with similar niches. *Vegetatio* 58: 29-55. <https://doi.org/10.1007/BF00044894>
- Shuai Y., Wang X.J., 2023. Niche and Interspecific Association of Dominant Tree Species in Spruce-Fir Mixed Forests in Northeast China. *Forests* 14(8): 1513. <https://doi.org/10.3390/fl14081513>
- Siebert F., 2012. A phytosociological synthesis of Mopanieveld vegetation at different spatial scales using various classification methods. North West Province, South Africa: North-West University.
- Small C.J., Mccarthy B.C., 2005. Relationship of understory diversity to soil nitrogen, topographic variation, and stand age in an eastern oak forest, USA. *Forest ecology and Management* 217(2-3): 229-243. <https://doi.org/10.1016/j.foreco.2005.06.004>
- Šmilauer P., Lepš J., 2014. *Multivariate analysis of ecological data using CANOCO 5*. Cambridge, United Kingdom: Cambridge University Press.
- Song N., Zhang J., 2018. Multivariate analysis of the endangered medicinal species *cercidiphyllum japonicum* communities in the Shennongjia Reserve, central China. *Cerne* 24: 180-189. <https://doi.org/10.1590/01047760201824032499>
- Soukhavong M., Yong L., Nanthavong K., Millet J., 2013. Investigation on species composition of plant community at Tad Xai at Phou Khao Khouay National Park, Lao PDR. *Our Nature* 11(1): 1-10. <https://doi.org/10.3126/on.v11i1.8237>
- Su S.J., Liu J.F., He Z.S., Zheng S.Q., Hong W., Xu D.W., 2015. Ecological species groups and interspecific

- association of dominant tree species in Daiyun Mountain National Nature Reserve. *Journal of Mountain Science* 12: 637-646. <https://doi.org/10.1007/s11629-013-2935-7>
- Van Der Maarel E., Sykes M.T., 1993. Small-scale plant species turnover in a limestone grassland: the carousel model and some comments on the niche concept. *Journal of vegetation science* 4(2): 179-188. <https://doi.org/10.2307/3236103>
- Vilches B., De Cáceres M., Sánchez-Mata D., Gavilán R.G., 2013. Indicator species of broad-leaved oak forests in the eastern Iberian Peninsula. *Ecological indicators* 26: 44-48. <https://doi.org/10.1016/j.ecolind.2012.10.022>
- Wang L.L., Cui X.Y., 2023. Niche of underground resources of *Pinus koraiensis* and common broadleaved tree species in broadleaved secondary forest of Xiaoxing'anling Mountains of northeastern China. *Journal of Beijing Forestry University* 44(12): 52-60. <http://j.bjfu.edu.cn/cn/article/doi/10.12171/j.1000-1522.20220082>
- Wiegand T., Gunatilleke S., Gunatilleke N., 2007. Species associations in a heterogeneous Sri Lankan dipterocarp forest. *The American Naturalist* 170(4): E77-E95. <https://doi.org/10.1086/521240>
- Xue W., Li C., Ai X., Yao L., Guo Q., Zhu J., 2023. Niche and interspecific association of dominant tree species in *Liriodendron chinense* natural forest. *Journal of Forest and Environment* 43(1): 26-34. <https://link.oversea.cnki.net/doi/10.13324/j.cnki.jfccf.2023.01.004>
- Yang Q.W., Liu S.J., Hu C.H., Lin Y.B., Zhang B., Luo M.L., Peng H.L., 2016. Ecological species groups and interspecific association of vegetation in natural recovery process at Xiejadian landslide after 2008 Wenchuan earthquake. *Journal of Mountain Science* 13: 1609-1620. <https://doi.org/10.1007/s11629-016-3807-8>